Effects of Prescribed Fire and Fire Surrogates on Saproxylic Coleoptera in the Southern Appalachians of North Carolina¹

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Abstract We examined the effects of forest management practices (prescribed burning, mechanical, and prescribed burn plus mechanical) on saproxylic forest Coleoptera in the southern Appalachian Mountains of North Carolina. During the 2-yr study, we captured 37,191 Coleoptera with baited multiple-funnel traps and pipe traps, comprising 20 families and 122 species that were used for our analysis. Saproxylic beetle numbers increased greatly from the first year to the second year on all treatments. Species richness and total abundance of Coleoptera were not significantly affected by the treatments, but several families (e.g., Elateridae, Cleridae, Trogositidae, Scolytidae) were significantly more abundant on treated plots. Abundances of many species, including various species of Scolytidae were significantly affected by the treatments. However, these scolytids (Hylastes salebrosus Eichoff, Ips grandicollis Eichoff, Xyloborinus saxeseni Ratzburg, Xyleborus sp., Xyleborus atratus Eichoff) did not respond in the same way to the treatments. Likewise, other Coleoptera such as Pityophagus sp. (Nitidulidae), Hylobius pales Herbst (Curculionidae), and Xylotrechus sagittatus Germar (Cerambycidae) also varied in their responses to the treatments. Species richness was not significantly different for the spring 2003 trapping seasons, but the fall 2003 sample had a higher number of species on mechanical shrub removal only and mechanical shrub removal plus prescribed burning plots compared with controls. Linear regression analysis suggests that increased dead wood caused by hot fires on mechanical plus burn and burn only treatments resulted in increases among various Coleoptera families and species. We saw no evidence that the treatments negatively impacted saproylic species and in most cases they benefited from the disturbances.

Key Words prescribed burn, forest management, Scolytidae, saproxylic insects, multiple funnel trap, coarse woody debris

Saproxylic beetles are a diverse insect group that is dependent on dead wood for food and habitat. They are the most abundant of all saproxylic invertebrates (Berg et al. 1994) and comprise more than 30% of the beetle fauna found within a forest (Speight 1989). Saproxylic beetles are an important component of a forest ecosystem because they contribute to wood decomposition and nutrient cycling and add organic matter into the soil (Grove 2002). Woody material is broken down by the beetles through tunneling and feeding action, and indirectly by facilitating bacterial and fungal growth that cause wood decay (Speight 1989). In addition, saproxylic insects are

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important components of the food web supporting a variety of invertebrate and vertebrate fauna (Tanner 1941, Harmon et al. 1986, Speight 1989, Hanula and Horn 2004).

Some saproxylic beetles (e.g., Scolytidae) are considered pests because they cause economic damage by killing trees or degrading wood quality. However, the few pest species readily recognized by landowners and researchers are greatly outnumbered by lesser known beetles that are valuable forest ecosystem components with little or no known negative economic impacts. Therefore, understanding how these unique Coleoptera respond to various forest management practices is important for conserving them. In many areas, forest management is the main influence affecting forest dynamics (Kuuluvainen 2002). The majority of saproxylic species are specialists and not very mobile, which makes them susceptible to habitat isolation resulting from large distances between pieces of dead wood (Nilsson and Baranowski 1997, Schiegg 2000). Forestry management practices often negatively impact many saproxylic beetle species (Niemela et al. 1996, Kuuluvainen 2002) because most of these practices decrease dead wood within forests (Ranius et al. 2005, Wikars et al. 2005) contributing to a loss of biodiversity (Fridman and Walheim 2000). Areas unaffected by such forestry practices have higher diversities of saproxylic organisms (Trave 2003). Thinning forests and removing dead and dying trees reduce needed resources for saproxylic species. However, species are not equally affected by reductions in dead wood so some species decline whereas others are unaffected (Kaila et al. 1994).

Due to long-term intensive forestry, many saproxylic beetles are endangered or extinct in parts of Europe (Kaila et al. 1994, Hammond et al. 2004, Wikars et al. 2005). Other areas of the world may begin to see decreases in saproxylic beetle diversity if intensive forest management is practiced over extensive areas. Most saproxylic invertebrate fauna are poorly known and potential impacts of various forestry practices are currently not well understood for many areas (Hammond et al. 2004), including the southeastern U.S. Therefore, we measured relative abundances and species richness of saproxylic beetle families and species after various forest management practices were applied in the southern Appalachian Mountains to determine how these practices might affect some early successional species.

Materials and Methods

This study was part of the National Fire and Fire Surrogate Study which is designed to examine the impacts of fuel reduction treatments on multiple components of forested ecosystems across the U.S. (Youngblood et al. 2005). We collected Coleoptera on the Green River Game Management Area, near Hendersonville, NC (Polk and Henderson counties) in the southern Appalachian Mountains. This forest is managed by the North Carolina Wildlife Resources Commission and encompasses 5,841 ha managed for game habitat and ecosystem restoration.

Twelve study sites were selected on the basis of size, stand age, cover type and management history. Each site was 14 ha in size to allow for a 10 ha measurement area and a buffer of at least 1 tree length (20 m) around the measurement area. All selected sites were judged to be in danger of uncharacteristically severe wildfire due to heavy fuel loads. None had been thinned during the past 10 yrs nor had any sites been burned (wild or prescribed) in at least 5 yrs. Stand ages varied from 80-120 yrs. Oaks dominated all sites including northern red oak (*Quercus rubra* L.), chestnut oak

(*Q. prinus* L.), white oak (*Q. alba* L.), and black oak (*Q. velutina* Lamarck). Other common species included pignut hickory (*Carya glabra* P.Mill.), mockernut hickory (*C. tomentosa* Poiret) and shortleaf pine (*Pinus echinata* P.Mill.). A thick shrub layer composed primarily of mountain laurel (*Kalmia latifolia* L.) and rhododendron (*Rhododendron maximum* L.), occurred on approx. one-half of the study area.

The study was a randomized complete block design consisting of 3 blocks of 4 treatments each. Treatments were applied to 10-ha plots and consisted of: (1) untreated, (2) dormant season burn, (3) mechanical, and (4) mechanical plus dormant season prescribed burn. The mechanical treatments consisted of chainsaw felling and bucking of the shrub understory, which was primarily composed of rhododendron, mountain laurel and small diameter trees (<7.5 cm). The felled material was left in place after cutting. The mechanical plus burn plots were treated the same way and later burned. One block was burned by hand ignition using spot fire and strip-headfire techniques. The other blocks were ignited by helicopter using a spot fire technique. Fire intensity was moderate to high with flame lengths of 1-2 m throughout the burn unit, but flames reached as high as 5 m in localized spots where topography or intersecting flame fronts contributed to erratic fire behavior. Shrubs were cut during the winter of 2001-2002, and plots that were burned were treated on 12 or 13 March 2002. Each plot was marked by grid points with 50 m between points to facilitate pre and posttreatment sampling.

Trap design has been shown to affect the number and species of beetles captured. We used 8-unit and 12-unit multiple funnel traps (PheroTech, Delta, BC), as well as modified pipe traps (D. Miller, unpubl. data) to capture flying saproxylic Coleoptera. Multiple funnel traps are effective for assessing abundance and diversity of Coleoptera (Chenier and Philogene 1989). Pipe traps work well for certain weevils and wood borers (Cerambycidae) and capture a variety of other beetles as well (D. Miller, pers. comm.). One of each trap type was used on each plot. The traps were suspended on a nylon string stretched between 2 trees. An ultra-high release α -pinene packet (PheroTech) and a 95% ethanol packet were placed within the 12-unit funnel trap and the pipe trap; whereas, only a 95% ethanol packet was placed in the 8-unit funnel trap. Collecting cups at the bottom of each trap were partially filled with propylene glycol to preserve captured insects. The pipe trap consisted of a solid, black, 1-m long, 15.2-cm diam. PVC pipe in place of the funnels and a large diameter collecting funnel (Miller, unpubl. data). Alpha-pinene is one of the common monoterpenes found in the resin of most pine species (Mirov 1961), whereas ethanol is a by-product of decomposition and a general attractant for a variety of bark and wood boring beetles (Fatzinger 1985, Fatzinger et al. 1987). Traps were placed 50 m apart near the center of each plot. In 2003, we trapped during the spring and fall for 10 wks each time, and in 2004 we trapped during the spring for 12 wks. A 2004 fall sample was not obtained because traps were damaged by Hurricane Ivan. During each trapping period, trap samples were gathered and the collecting cups were refilled with propylene glycol every 3-4 wks. Samples were stored in 70% ethyl alcohol until they were identified.

The density of trees remaining on the plots was estimated by measuring tree basal area (Avery 1975) on ten 0.2-ha subplots within each 10-ha treatment plot. Basal area was measured in 2001 (pretreatment) and in 2004-2005 (posttreatment). Because basal area should increase with time in undisturbed stands, we used change in basal area (posttreatment minus pretreatment basal area) as an indicator of treatment effects on dominant trees. Decreases in basal area were the result of treatment-related tree mortality.

Data were analyzed using PROC GLM (SAS 1985) to conduct two-way ANOVAs with replications and treatments as dependent variables, and the various families, genera or species of Coleoptera as independent variables. The Ryan-Einot-Gabriel-Welsch multiple range test was used to determine differences in relative abundances and diversities of Coleoptera between treatments. Square-root transformation was used to assure normality and homogeneity of variance. For statistical analysis we compared the numbers caught and species richness of spring 2003, fall 2003, spring 2004 separately, and all data together. We used the GLM procedure to calculate simple linear regressions of families and species versus change in basal area.

Results

During the 2 yrs of sampling, we captured 37,191 saproxylic Coleoptera and associated predators comprising 20 families and 122 species (Table 1). Overall, species richness (Fig. 1) and total numbers of saproxylic Coleoptera (Fig. 2) were not significantly different among treatments, but differences in abundance were observed at the family and species level (Table 2). For example, Elateridae were significantly higher in samples from mechanical shrub control plus burn treatments than controls in spring 2003 and in spring 2004 the mechanical shrub control plus burn and mechanical shrub control only plots had more elaterids than burn only and control plots (Table 2). Trogositidae were captured in higher numbers on mechanical shrub control only treatment plots compared with burn only plots in spring 2003 but not in other trapping periods. Cleridae abundance in spring 2004 was significantly higher on mechanical shrub control plus burn treatments compared with all other treatments. Overall, Scolytidae numbers increased greatly from the first year to the second in all treatments (Fig. 3). In fall 2003, Scolytidae were captured in higher numbers on mechanical shrub control plus burn treatment plots compared with other treatments, but they did not differ significantly during other sample periods (Fig. 3, Table 2).

In 2003, 4 species of Scolytidae were significantly different in abundance among treatments (Table 1). Hylastes salebrosus Eichoff (P = 0.06) were caught in higher numbers on the burn-only plots compared with control plots in spring 2003, but in fall 2003 they were captured in higher numbers (P = 0.07) on mechanical shrub control only treatment plots compared with the other 3 treatments. Ips grandicollis Eichoff were caught in higher numbers on mechanical shrub control only treatment plots in spring 2003 but in fall 2003 more were caught on the mechanical shrub control plus burn treatment. In spring 2004, I. grandicollis numbers were higher on the mechanical shrub control plus burn and mechanical shrub control only treatment plots than on the burn-only plots and the mechanical shrub control plus burn plots had more than the controls. In spring 2003, Xyloborinus saxeseni Ratzburg and Dryoxylon onoharaensum Murayama were more abundant on mechanical shrub control treatments compared with burn only and mechanical shrub control plus burn. In the spring 2004, Dryoxylon onoharaensum followed the same pattern but X. saxeseni were more abundant on the mechanical shrub control plus burn treatment compared with the control.

Species richness was not significantly different for either spring trapping season (Fig. 4). The fall 2003 sample had the lowest species richness of the 3 trapping periods, but the number of species was significantly higher on mechanical shrub control only and mechanical shrub control plus burn plots compared with controls

Table 1. Coleoptera genera and species and total numbers captured with multiple funnel traps on fire and fire surrogate treatments on the Green River Game Management Area near Hendersonville, NC

Family	Genus/Species	Total Number Captured
Anobiidae	Hadrobregmus sp.	58
	Hemicoelus sp.	68
	Trichodesma sp.	2
Anthribidae	sp. 1	38
	sp. 2	80
	Eurymycter fasciatus	7
	Toxonotus sp.	2
Bostrichidae	Amphicerus bicaudatus	1
	sp. 1	8
	Xylobiops sp. 1	8
Brentidae	Arrhenodes minutus	5
Buprestidae	Acmaeodera sp.	1
	Agrilis <i>sp. 1</i>	2
	Agrilis sp. 2	1
	Buprestis salisburyensis	1
	Buprestis sp. 1	9
Cerambycidae	Acanthocinus obseletus	11
	Acanthocinus pusillus	3
	Acanthocinus sp.	1
	Aneflomorpha sp.	19
	Arhopalus rusticus	3
	Brachysomida bivittata	5
	Clytus marginicollis	4
	Clytus ruricola	351
	Cyrtophorus verrurosus	90
	Gaurotes cyanipennis	17
	Judolia cordifera	6
	Knulliana cincta	3
	Leptostylus sp.	7
	Leptura plebeja	1
	Leptura sp. 1	1
	Megacyllene caryae	3

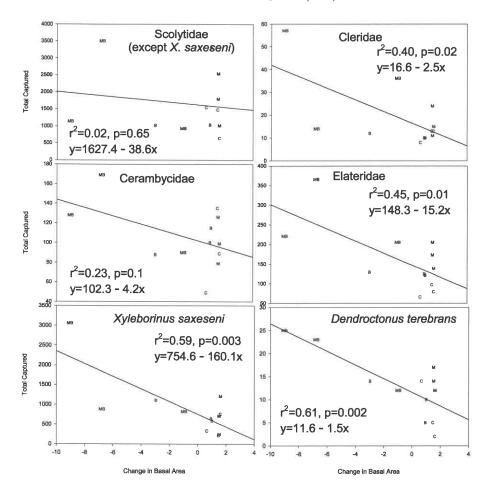


Fig. 5. Linear regressions of various saproxylic Coleoptera captured in funnel traps and change in basal area of 10 ha plots treated with various combinations of prescribed fire and mechanical brush removal. Negative numbers indicate a post treatment decrease in basal area (more dead wood on plots). MB = mechanical plus burn, M = mechanical, B = burn, C = control

tensive areas that experienced low to high wildfire severity, so these results might indicate that this beetle is less likely to occur in burned areas. However, in spring 2003 it was more abundant on burned areas than on the controls, and Sullivan et al. (2003) found *H. salebrosus* was attracted to small burned plots, so factors other than fire alone likely affected *H. salebrosus* populations in this study. Likewise, *D. terebrans* abundance increased as basal area decreased. In all plots with high *D. terebrans* populations, fire was the cause of the tree mortality in this study, but the beetles were either attracted to the areas by the dead trees or used them for brood production. In comparison, Hanula et al. (2002) caught fewer *D. terebrans* in recently burned areas

that were part of a 15,000 ha wildfire. Our lack of capture success for *lps* species could be a bias in the funnel traps because Smith et al. (1993) showed that some species of *lps*, such as *l. avulsus* Eichoff and *l. calligraphus* Germar were not highly attracted to ethanol or turpentine.

Increased abundance of X. saxeseni in spring 2003 may have been a result of fire reducing the available habitat. This species was probably able to use the cut rhododendron and mountain laurel because it has been found feeding on nearly all genera of deciduous trees (Furniss and Johnson 2002). The mechanical shrub control only treatment plots had an abundance of freshly cut wood in 2003, and the controls had normal background levels of newly dead material. On the other hand, prescribed burns were applied just before our spring 2003 trapping period and reduced the amount of wood from the cut shrubs and small trees or may have rendered it unacceptable as host material. By 2004 beetles were likely emerging from the fire-killed trees on the mechanical plus burn plots. Also, our traps may have competed with the scents released by the freshly cut wood and recently fire-killed trees in 2003, but by 2004 this competition should have ceased. Thus, newly-emerged brood and reduced competition between traps and freshly cut wood might explain the large numbers of X. saxeseni captured in 2004 compared with 2003. Because we do not know host plants of Dryoxylon onoharaensum, increased abundance also could be due to increased amounts of dead rhododendron and mountain laurel. The higher elaterid abundance in 2004 probably resulted from the increased dead wood availability or emergence of brood from the previous year.

An introduced species that has a wide host range, X. crassiusculus Motschulsky was captured in large numbers during this study. First discovered in South Carolina in 1974, it is considered a major pest in nurseries (Oliver and Mannion 2001) and a potential problem in forests because they can attack healthy trees (Atkinson et al. 1988). The spring 2003 sample contained few X. crassiusculus but captures increased in the fall 2003 and by spring 2004 our samples contained large numbers. Due to the likely increase in scents from cut and burned wood, our traps may have competed with these odors in 2003; whereas, in 2004 the scents had dissipated and our traps were much more effective. The increase in cut wood in 2003 would have allowed for more beetles to breed and possibly resulted in a large emergence by 2004, which could also explain the rise in numbers of X. crassiusculus and other beetles in 2004. However, captures of this species were not correlated with decreasing basal area resulting from fire-caused tree mortality, nor did any of the treatments result in higher numbers, so the reasons for the population increase are unclear. Like X. crassiusculus, X. saxeseni, X. atratus, and Dryoxylon onoharaensum are nonnative species. Because X. crassiusculus, X. saxeseni, and D. onoharaensum comprised the large majority (~80%) of our captures of Scolytidae, it raises the question of whether these species are displacing native insects.

Pityophagus sp. (Nitidulidae) was captured on burn-only and controls in significantly higher numbers compared with mechanical shrub control plus burn plots in spring 2003. It is unknown why its abundances would be affected by these treatments. This genus is predatory on bark beetles and has only rarely been found in the southeastern United States.

Because we used 3 traps per plot and our plots were relatively close to each other, the proximity of the plots and large amount of α -pinene and ethanol released from our traps may have attracted beetles from outside the treatment plots or they may have been able to move from one treatment plot to the next. Future studies may want to

consider using only 1 funnel trap per plot, using larger plots, or passive traps like flight intercept traps.

Beetles varied in their responses to the treatments but, in general, treatments that created the most tree mortality resulted in greater numbers of saproxylic species. The reintroduction of fire into forests where it has been excluded may initially benefit saproxylic organisms. Our samples included primarily early successional species and dead wood created by these treatments should be an important resource for later successional species for some time. However, because the stands were "thinned" by fire it is unlikely that much additional tree mortality will occur in the near future because the fire reduced competition and tree vigor should improve. Future research should focus on long-term changes among Coleoptera assemblages, the effects of intermittent versus continuous inputs of dead wood on saproxylic species, and the impacts of nonnative species on native saproxylic fauna.

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